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Concentrations of Polychlorinated Dibenzo-p-Dioxins (PCDD) and Dibenzofurans (PCDF) in Urban Runoff and Household Wastewaters

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Abstract

Urban runoff and household wastewater samples were analysed for PCDD/F. The concentrations in street runoff lay between 1 and 11 pg TEQ/L. A roof runoff sample was less contaminated (< 1.7 pg TEQ/L), with levels at most a factor of 2 higher than in a parallel rain water sample. Household wastewater showed higher PCDD/F concentrations of up to 14 pg TEQ/L. The homolog profile agreed well with the profile found in sewage sludge samples. Washing machine effluent is a major source of PCDD/F in household wastewater. PCDD/F were also quantified in shower water.

Keywords: PCDD/F, Street Runoff, Roof Runoff, Household Wastewater, Sewage Sludge, Shower Water, Washing Machine Effluent

Introduction

Polychlorinated dibenzo-p-dioxins and dibenzofurans (PCDD/F) are persistent, lipophilic and, depending on the chlorine substitution, toxic compounds that accumulate in different environmental compartments. An ubiquitous contamination of sewage sludge with these compounds has been reported [1,2,3], and it has been shown that the agricultural use of sewage sludge as a fertiliser can cause an accumulation of PCDD/F in soils [4]. Very similar PCDD/F homolog profiles have been observed in sewage sludge samples collected in municipal wastewater treatment plants (MWTP) across Germany [5]. The similarity between this typical sludge profile and the profile measured in technical pentachlorophenol (PCP) led to the hypothesis that the use of technical products containing PCP in different industries was the main source of the PCDD/F in sewage sludge [6]. However, PCDD/F contamination displaying the same profile was also found in sewage sludge with no industrial inputs and in isolated household wastewater treatment systems [5,7]. Attempts have been made to resolve this contradiction and identify the PCDD/F sources in detail [8,9], with limited success. In an earlier paper we provided evidence indicating that household wastewater and not surface runoff contributes the majority of the PCDD/F in sewage sludge [10]. We more recently proposed that a large portion of the PCDD/F originates in textiles, and that the PCDD/F are washed out of the textiles in the washing machine as well as being washed off of human skin during bathing [11]. In this paper we provide full isomer specific concentration data on a range of the samples analysed in this work to assist in future efforts to quantify PCDD/F fluxes in municipal wastewater. Results are presented for urban street runoff, storm water sediment trap sediments, sewer sediments and household wastewater collected in Bayreuth, a small city of 70,000 in northeastern Bayaria, as well as for washing machine effluent and shower water.

Analytical Methods

The aqueous and particulate phases of runoff water and household wastewater were separated by filtration. The filtrate was extracted three times with toluene (20% of the sample volume) in a separatory funnel. The filters were extracted in toluene for 24 hours using a Soxhlet apparatus as were the sediment samples following drying at 75°C. An internal standard mixture containing 11 ¹³C-labelled and one ³⁷Cl-labelled 2,3,7,8-substituted PCDD/F congeners in toluene was added to either the solvent or the sample prior to extraction of all samples.

A solid phase extraction (SPE) method was developed for extracting the washing machine effluent and shower water. PCDD/F from unfiltered samples of up to 2000 mL were enriched on 26 mm cartridges containing octadecyl silane (C₁₈) bonded to a silica substrate (VARIAN MEGA BOND ELUT; 10 g sorbent). The sorbent was conditioned with 20 mL n-hexane, 20 mL dichloromethane, 20 mL methanol and 300 mL aqua bidest under low vacuum. The samples were drawn through the C₁₈-sorbent at a maximum flow rate of 50 mL/min by applying a mild vacuum. After the enrichment the remaining water on the sorbent was removed by sucking air through for 5 minutes and subsequently allowing the cartridge to stand for 2 days at room temperature. The internal standard mixture was then added to the top of the cartridge and the PCDD/F were eluted with 50 mL of toluene. The solvent was transferred directly to the first column used for the sample cleanup. The extraction efficiency was tested with five laundry wastewater samples which were divided in two equal parts. One half was extracted with the SPE-method, while the extraction of the parallel sample was done by liquid/liquid extraction with toluene. The PCDD/F concentrations were compared for each homolog of the five sample pairs. In comparison to the liquid/liquid extraction with toluene (=100%) the average SPE efficiency lay between 97% and 130% for all PCDD/F homologs. The SPE method saved time and required only 5% of the solvent used in liquid/liquid extraction.

The cleanup and HRGC/HRMS analysis of all samples were as described in [12]. The 2,3,7,8-Cl₄DD toxicity equivalents (TEQ values) were calculated using the NATO-CCMS factors [13].

Results and Discussion

Runoff Water

Street runoff samples were collected at a busy intersection in Bayreuth (more than 16,000 vehicles per day) at the beginning and end of two rain events in July 1991. About 4 L of runoff water were sampled with an automated device developed by the Hydrology Department of the University of Bayreuth and described elsewhere [14]. This device was constructed without a sediment trap to avoid loss of sediment material during sampling. Litres 3 - 6 of a runoff event sampled as "first flush", while the sixth last to third last litres were sampled as "last flush".

The first rain event on July 14 lasted eight hours. The rain intensity was relatively constant over this period (0.3 mm/h - 0.55 mm/h). Table 1 gives the PCDD/F-concentrations of the samples from the beginning and end of this event. On July 26 two more samples were taken at the beginning and end of a runoff event caused by a thunderstorm. In contrast to the first rain event the intensity was very high at the beginning of the storm (3.5 mm/h) and decreased rapidly to zero after 30 minutes. The PCDD/F-concentrations of the second rain event are also shown in Table 1. The concentration values include the aqueous and the particulate phases but in all four cases more than 99.9% of the PCDD/F were adsorbed to the particulate matter.

The PCDD/F concentrations in the runoff from 14.07.91 were approximately constant in both samples (< 2.5 pg TEO/L in the first flush; < 3.5 pg TEQ/L in the last flush). Only Cl₈DD showed a slight increase at the end of the first event. During the second event the PCDD/F levels decreased by a factor of 5 to 12 (10.2 pg TEQ/L in the first flush; < 1.9 pg TEQ/L in the last flush). This was likely due to changes in the rain intensity and its effect on the cleansing of the street surface. Because PCDD/F are almost completely adsorbed to particulate matter the contaminant flux is determined by the concentration on the particles and the amount of particulate matter The PCDD/F flux during the first event was likely quite constant since the resuspended by the runoff. concentration of particulate matter in the runoff was relatively constant. Some 0.27 g/L were present in the first flush and 0.11g/L in the last flush, where the difference was caused by the absence of sandy material in the last sample. The fine material that is responsible for most of the PCDD/F flux was present in similar quantities in both samples. During the second rain event the flux of particulate matter decreased over time (0.05 g/L in the last flush as opposed to 1.1 g/L in the first flush). This can be attributed to either the decreasing rain intensity and/or a complete cleansing of particulate material from the street during the initial very intensive rain. More detailed studies of other organic compounds (i.e. PAH) with physical-chemical properties similar to PCDD/F showed a strong correlation between the sediment flux and the pollutant flux in street runoff [14,15,16].

Tab. 1: PCDD/F concentrations in urban street runoff from 14.07.91 and 26.07.91 (pg/L)

Street Runoff:	First Flush, 14.07.91	Last Flush, 14.07.91	First Flush, 26.07.91	Last Flush, 26.07.91
Σ Cl ₄ DD	3.5	3	15	1.3
Σ Cl ₅ DD	14	13	23	6.6
Σ Cl ₆ DD	29	25	58	13
Σ Cl ₇ DD	52	60	125	20
Cl ₈ DD	120	200	420	60
Σ Cl ₄ DF	20	21	72	16
Σ Cl ₅ DF	16	15	49	11
Σ Cl ₆ DF	23	20	69	12
Σ Cl ₇ DF	22	22	73	11
Cl ₈ DF	: 13	16	52	4.3
2,3,7,8-Cl ₄ DD	< 0.5	< 0.5	< 0.5	< 0.2
1,2,3,7,8-Cl ₅ DD	< 0.5	< 0.5	< 0.5	< 0.2
1,2,3,4,7,8-Cl ₆ DD	< 0.5	1.2	< 0.5	< 0.2
1,2,3,6,7,8-Cl ₆ DD	< 0.5	2.9	19	< 0.2
1,2,3,7,8,9-Cl ₆ DD	< 0.5	< 0.5	< 0.5	< 0.2
1,2,3,4,6,7,8-Cl ₇ DD	24	22	68	7
2,3,7,8-Cl ₄ DF	1.9	1.5	8.2	1.2
1,2,3,4/7,8-Cl ₅ DF	< 0.5	1.1	5.7	< 0.2
2,3,4,7,8-Cl ₅ DF	< 0.5	2.0	6.3	1.2
1,2,3,4,7,8/9-Cl ₆ DF	< 0.5	1.2	5.8	0.9
1,2,3,6,7,8-Cl ₆ DF	< 0.5	2.9	5.6	0.9
1,2,3,7,8,9-Cl ₆ DF	< 0.5	< 0.5	< 0.5	< 0.5
2,3,4,6,7,8-Cl ₆ DF	< 2.5	1.8	12	0.6
1,2,3,4,6,7,8-Cl ₇ DF	14	14	48	3.2
1,2,3,4,7,8,9-Cl ₇ DF	1.4	1.0	5.5	0.22
TEQ	< 2.5*/ 1,0**	< 3.5*/ 2,9**	< 11*/10**	< 1.9*/ 1.2**

[&]quot;<" concentrations included in calculating the TEQs; ** "<" concentrations not included in calculating of TEQs

On 26.07.91 rain water was sampled simultaneously to the street runoff. The rain water contained 21 pg/L Cl₈DD and less of the other PCDD/F-homologs [17]. The PCDD/F-concentrations in the first flush of the street runoff lay up to 20 times higher than in the rain water. In the last flush the difference was only a factor of 2. This indicates that other sources besides rain are responsible for most of the PCDD/F in street runoff. Car engines have been identified as a source of PCDD/F, especially car engines which use leaded fuel [18]. It is likely that residues from car and truck exhaust contribute to the PCDD/F load in street runoff.

Sediment material (< 0.063 mm) from different storm water sediment traps was analysed for PCDD/F. Three locations in Bayreuth with different traffic densities were sampled: Albrecht-Dürer-Str. with >20,000 cars per day; Königsallee with 16,000 cars per day, and Hasenweg with 2,000 cars per day. The concentrations are presented in Table 2.

Tab. 2: PCDD/F in sediments from storm water sediment traps (ng/kg)

Sediments (<0,063mm):	Albrecht-Dürer-Straße	Königsallee	Hasenweg
Σ Cl ₄ DD	170	38	26
Σ Cl ₅ DD	210	57	33
Σ Cl ₆ DD	340	110	87
Σ Cl ₇ DD	390	320	240
Cl ₈ DD	860	880	650
Σ Cl ₄ DF	460	100	88
Σ Cl ₅ DF	430	83	54
Σ Cl ₆ DF	330	88	53
Σ Cl ₇ DF	200	110	68
Cl ₈ DF	110	73	46
2,3,7,8-Cl ₄ DD	4.9	< 0.1	< 0.1
1,2,3,7,8-Cl ₅ DD	. 12	4.8	3.2
1,2,3,4,7,8-Cl ₆ DD	23	4.0	2.6
1,2,3,6,7,8-Cl ₆ DD	45	12	5.0
1,2,3,7,8,9-Cl ₆ DD	19	6.5	4.2
1,2,3,4,6,7,8-Cl ₇ DD	150	150	110
. 2,3,7,8-Cl ₄ DF	9.0	8.2	7.3
1,2,3,4/7,8-Cl ₅ DF	5.2	8.0	5.0
2,3,4,7,8-CI ₅ DF	7.0	8.3	5.3
1,2,3,4,7,8/9-Cl ₆ DF	3.5	8.1	5.3
1,2,3,6,7,8-Cl ₆ DF	3.1	6.1	3.2
1,2,3,7,8,9-Cl ₆ DF	2.2	1.4	< 0.5
2,3,4,6,7,8-Cl ₆ DF	1.8	6.2	4.2
1,2,3,4,6,7,8-Cl ₇ DF	130	64	39
1,2,3,4,7,8,9-Cl ₇ DF	12	8.5	4.1
TEQ	29	15	10

The results support the hypothesis stated above. The PCDD/F concentrations were highest in sediment material from the Albrecht-Dürer-Straße, the street with the highest traffic density. The homolog profile in this sample showed a larger portion of lower chlorinated PCDD and PCDF in comparison to other samples, an observation that was also made by Schwind for various samples of automobile exhaust [18].

Two samples of roof runoff were collected at the end of October 1991 and analysed for PCDD/F. The concentrations (TEQ: < 1.7 pg/L; Cl₈DD: 50-100 pg/L) were at most a factor of 2 higher than in a parallel rain sample. The homolog profile of roof runoff and rain water agreed well.

The surface areas covered by streets (480 ha) and roofs (500 ha) are nearly equal in Bayreuth. Hence, the higher concentrations measured in street runoff indicate that this is a more important source of PCDD/F to the MWTP than roof runoff.

Sediments (< 0.25 mm) from different parts of the municipal sewer system in Bayreuth were also analysed for PCDD/F. Among the samples were one taken from a sewer servicing only households and surface runoff, one servicing households, surface runoff and industry, one servicing only street runoff, and one from a rain water retention basin. The results are presented in Table 3.

Tab. 3: PCDD/F-concentrations in different sediments (<0.25 mm) from sewer systems (ng/kg)

Sewer Sediment Samples:	Industry, Runoff	Household, Runoff	Runoff	Runoff Basin Sediments
Σ Cl ₄ DD	5.9	1.5	7.7	10
Σ Cl ₅ DD	11	6.7	14	14
Σ Cl ₆ DD	19	33	32	39
Σ Cl ₇ DD	45	130	82	120
Cl ₈ DD	220	770	270	410
Σ Cl ₄ DF	. 28	21	28	47
Σ Cl ₅ DF	35	23	26	38
Σ Cl ₆ DF	36	14	30	41
Σ Cl ₇ DF	27	11	25	53
Cl ₈ DF	18	13	18	36
2,3,7,8-Cl ₄ DD	< 0.1	< 0.1	0.17	0.49
1,2,3,7,8-CI ₅ DD	< 0.1	< 0.1	0.80	1.2
1,2,3,4,7,8-Cl ₆ DD	0.92	1.9	0.82	2.5
1,2,3,6,7,8-Cl ₆ DD	2.2	4.3	3.1	5.9
1,2,3,7,8,9-Cl ₆ DD	1.7	2.4	2.0	4.0
1,2,3,4,6,7,8-Cl ₇ DD	24	40	42	60
2,3,7,8-Cl ₄ DF	2.4	4.4	2.0	4.8
1,2,3,4/7,8-Cl ₅ DF	6.6	4.0	0.43	4.1
2,3,4,7,8-Cl ₅ DF	3.0	3.2	1.6	3.6
1,2,3,4,7,8/9-Cl ₆ DF	4.4	5.5	1.8	4.6
1,2,3,6,7,8-Cl ₆ DF	4.6	2.5	2.9	5.3
1,2,3,7,8,9-Cl ₆ DF	< 0.5	<0.5	<0.5	<0.5
2,3,4,6,7,8-Cl ₆ DF	3.0	1.6	1.9	4.0
1,2,3,4,6,7,8-Cl ₇ DF	17	6,6	14	29
1,2,3,4,7,8,9-Cl ₇ DF	1.9	0.77	1.0	1.9
TEQ	4.4	5.3	3.7	7.6

The highest concentrations of the higher chlorinated PCDD/F were found in sewers which carried primarily household wastewater and occasionally runoff water but without any industrial influence. The homolog profile agreed well with those in household wastewater and sewage sludge [11]. In contrast, higher concentrations of the lower chlorinated PCDD/F were found in the sewers with no contributions from household wastewater. The profiles from these latter samples were similar to the atmospheric patterns found in the surface runoff, but distinctly different from the pattern observed in sewage sludge. This was one of several pieces of evidence suggesting that household wastewater was an important source of PCDD/F in sewage sludge.

The TEQ-concentrations found in sewer sediments were distinctly lower than those found in sewage sludge from the MWTP in Bayreuth. During a sampling period in 1991 TEQ-concentration between 16 pg/g and 60 pg/g were detected in more than 25 sludge samples.

Household Wastewater

Household wastewater samples were collected on two different days (a Tuesday and a Friday) at an apartment building in Bayreuth (FRG). The apartment building had approximately 750 inhabitants and no commercial space. The building drainage for roof runoff and household wastewater were led through separate piping systems which made it possible to collect samples containing only household wastewater. The samples were collected with a ladle as the water fell over a weir. This allowed a representative sampling of the wastewater including the solid material. One ladle was collected every 5 minutes. The water collected over one hour was pooled to give a 3-4 L sample. The results are given in Table 4 along with the water quantity used in the building during the sampling periods and the total amount of water used over the day.

Table 4: PCDD/F in household wastewater (pg/L)

Household Wastewater:	Friday 15.11.91	Friday 15.11.91	Friday 15.11.91	Tuesday 31.03.92	Tuesday 31.03.92	Tuesday 31.03.92	Tuesday 31.03.92	Tuesday 31.03.92
Sampling time	7:00-8:00	11:00-12:00	15:00-16:00	7:00-7:30	9:20-9:40	11:00-11:30	14:50-15:05	21:50-22:10
Water quantity per hour	5.75 m³	4.17 m³	3.99 m³	6.60 m³	5.04 m³	3.33 m ³	2.97 m³	1.00 m³
Water quantity per day		80.3 m ³		84.0 m³				
Σ Cl₄DD	2.4	11	6.6	1.4	4.9	6.0	1,9	1.2
Σ CI ₅ DD	. 15	46	89	4.8	18	31	11	5.6
Σ Cl ₆ DD	11	110	130	6.3	23	36	12	8.8
Σ Cl ₇ DD	62	1100	920	15	97	170	28	24
Cl ₈ DD	360	4500	2600	100	690	840	170	100
Σ Cl ₄ DF	7.1.	26	19	5,8	16	18	9.7	7.9
Σ Cl ₅ DF	4.5	20	11	3.2	11	13	6.7	3.2
Σ Cl ₆ DF	5.3	30	15	2.9	10	15	6.7	2.8
Σ Cl ₇ DF	5.5	49	25	3.9	17	27	9.3	5.0
Cl ₈ DF	6.0	40	24	3.5	12	23	7.6	6.0
2,3,7,8-Cl ₄ DD	< 0.1	< 0.1	< 0.1	< 0.1	< 0.3	< 0.3	< 0.3	< 0.2
1,2,3,7,8-Cl ₅ DD	< 0.1	0:38	0.22	0.29	0.15	0.24	0.32	0.10
1,2,3,4,7,8-Cl ₆ DD	< 0.1	0.80	0.20	0.30	< 0.1	< 0.1	< 0.1	< 0.1
1,2,3,6,7,8-Cl ₆ DD	1.1	9.4	14	2.1	3.2	1.7	1.7	1.0
1,2,3,7,8,9-Cl ₆ DD	0.56	3.1	3.6	0.72	1.3	0.68	0.95	0.45
1,2,3,4,6,7,8-Cl ₇ DD	31	550	600	8.7	53	94	17	13
2,3,7,8-Cl ₄ DF	0.39	1.4	0.83	1.4	1.9	1.6	3.3	0.92
1,2,3,4/7,8-Cl ₅ DF	< 0.1	0.63	0.54	<0.1	0.45	0.56	0.59	< 0.1
2,3,4,7,8-Cl ₅ DF	0.46	1.6	1.1	0.83	0.78	0.61	0.36	0.10
1,2,3,4,7,8/9-Cl ₆ DF	0.50	1.3	1.0	0.59	1.4	< 0.1	< 0.1	< 0.1
1,2,3,6,7,8-Cl ₆ DF	0.54	4.2	2.5	0.39	0.93	< 0.1	< 0.1	< 0.1
1,2,3,7,8,9-Cl ₆ DF	< 0.1	0.83	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2,3,4,6,7,8-Cl ₆ DF	0.51	1.5	1.2	< 0.34	1.0	< 0.1	< 0.1	< 0.1
1,2,3,4,6,7,8-Cl ₇ DF	0.29	28	12	2.3	10	16	6.2	3.1
1,2,3,4,7,8,9-Cl ₇ DF	0.17	0.96	0.66	0.10	0.40	0.61	0.26	0.15
TEQ	1.4	14	12	1.7	3.1	3.3	1.8	0.81

The PCDD/F concentrations found in the household wastewater differed significantly in the course of one day and also between the two days. The highest concentrations were found during daytime. In the morning and also in the evening the PCDD/F-concentrations were much lower. The concentrations were up to 10 times higher than in the runoff samples.

An attempt at a crude mass balance with these data was made in another paper [10]. The flux of the PCDD/F in TEQs into the Bayreuth MWTP was estimated to be 8 times higher from household wastewater than from surface runoff. Also, the homolog profile differed between the two kinds of samples. The profile in household wastewater agreed well with the profile in sewage sludge whereas the profile in runoff samples was distinctly different (see above). The daily cycle in the household wastewater concentrations and observations of wastewater colour and consistency led us to suspect washing effluent as a possible PCDD/F source.

Washing Machine Effluent

Following up on the results of the last section, washing machine effluent samples were analysed for PCDD/F. Washing machine effluent was collected in different households and from different kinds of clothing. All wastewater from one washing cycle was collected in a big tub and mixed well. For each analysis 1-2 L of effluent were transferred to a glass jar and transported to the lab. A method blank was obtained by running the washing machine through its cycle with detergent but without clothes. The PCDD/F concentrations in the blank were at least a factor of 100 lower than those present in the samples with laundry. Hence, neither the detergent nor the machine itself nor the sampling procedure was the source of the PCDD/F.

The washing temperature varied depending on the type of laundry: white clothes were washed at 95°C, baby wash and dark coloured clothes at 60°C, and bright coloured clothes at 40°C. Table 5 gives the PCDD/F concentrations found in the different washing machine effluent samples. The PCDD/F flux per load of wash can be easily calculated by multiplying the volume of water by the concentration.

All four washing machine effluent samples had the same homolog profile that was found in household wastewater and sewage sludge [10,11]. The concentrations ranged up to 25 pg TEQ/L and varied by only a factor of 2 in the 4 samples. The homolog concentrations in the washing machine wastewater were very similar to those measured in the most contaminated household wastewater samples.

In further studies it was shown that textiles were contaminated with PCDD/F [11,12,19]. In order to examine the fate of PCDD/F on textiles during washing an experiment was conducted in which two mildly contaminated cotton T-shirts containing altogether 280 ng Σ PCDD/F (620 pg TEQ) were washed together with two uncontaminated cotton undershirts (1.5 ng Σ PCDD/F or 8 pg TEQ on both undershirts). After washing it was found that 16% of the PCDD/F (19% of the TEQ) originally present in the contaminated T-shirts was in the washing machine effluent and 7% (8% of the TEQ) in the previously uncontaminated undershirts [11]. Another experiment under somewhat different conditions (an analysed T-shirt with 3 kg common dirty wash) gave similar results: 35% of the PCDD/F were removed from the T-shirt during washing. This demonstrates that PCDD/F are not tightly bound to the fabric and washing out of these compounds does occur.

Tab. 5: PCDD/F-concentrations in washing machine effluent (pg/L)

Samples:	Dark coloured cloth	White cloth	Bright colored cloth	Baby cloth
Water quantity per wash:	821	571	671	861
Textiles per wash:	4 kg	2 kg	3.2 kg	3 kg
Σ Cl ₄ DD	30	40	31	21
Σ Cl ₅ DD	110	80	130	60
Σ Cl ₆ DD	330	240	530	170
Σ Cl ₇ DD	1100	630	1600	990
Cl ₈ DD	4100	2300	3600	4400
ΣCl_4DF	81	120	97	71
Σ Cl ₅ DF	67	150	120	100
Σ Cl ₆ DF	160	120	190	310
Σ CI ₇ DF	500	180	470	510
Cl ₈ DF	380	160	300	430
2,3,7,8-Cl ₄ DD	< 1	1.6	1.0	< 0.5
1,2,3,7,8-Cl ₅ DD	1.9	1.7	1.9	1.7
1,2,3,4,7,8-Cl ₆ DD		< 0.5	3.2	< 0.5
1,2,3,6,7,8-Cl ₆ DD		13	24	23
1,2,3,7,8,9-Cl ₆ DD		6.9	9.2	6.8
1,2,3,4,6,7,8-Cl ₇ DD	540	270	600	570
2,3,7,8-Cl ₄ DF	7.4	8.3	5.9	4.3
1,2,3,4/7,8-Cl ₅ DF	2.1	5.7	3.1	3,3
2,3,4,7,8-Cl ₅ DF	4.2	4.9	4.6	9.8
1,2,3,4,7,8/9-Cl ₆ DF	4.9	4.2	4.2	8.5
1,2,3,6,7,8-Cl ₆ DF	6.5	6.1	7.9	7.5
1,2,3,7,8,9-Cl ₆ DF	< 0.5	< 0.5	< 0.5	< 0.5
2,3,4,6,7,8-Cl ₆ DF	12	8.2	9.7	5.8
1,2,3,4,6,7,8-Cl ₇ DF	330	120	300	200
1,2,3,4,7,8,9-Cl ₇ DF	3.8	4.5	2.1	8.8
TEQ	25	17	24	25

Shower Water

Shower and bath water represent another potential source of PCDD/F to household wastewater. In average of Germany 30% (= 43 L) of the total daily water consumption is used for this purpose [20]. It was demonstrated that PCDD/F can transfer from contaminated textiles to the skin's surface [11], and it was thought that these PCDD/F on the skin's surface might be removed during washing and contribute to the PCDD/F levels in household wastewater. In order to test this hypothesis the shower water from 5 individuals was analysed. As the concentrations depend on the water quantity used for each shower, the results in Table 6 are given in pg/shower and not pg/L.

The PCDD/F-quantities in the shower water differed by nearly two orders of magnitude. This variability can be explained by the variability in textile contamination and hence in the expected skin contamination. The PCDD/F-homolog profile agreed well with typical profiles in textiles [12,19] and was similar to the profile found in household wastewater (→ Table 3) and in sewage sludge. These results show that wash water contributes to the PCDD/F load in household wastewater and are consistent with the hypothesis that skin is the source.

Tab. 6: PCDD/F-quantities in shower water (pg)

Shower water: (water quantity)	Sample 1 (13 litres)	Sample 2 (25 litres)	Sample 3 (19 litres)	Sample 4 (7.5 litres)	Sample 5 (5 litres)
Σ Cl ₄ DD	62	80	70	26	60
Σ Cl ₅ DD	83	200	150	71	84
Σ Cl ₆ DD	640	1700	510	290	280
Σ Cl ₇ DD	8500	5000	1500	980	60
Cl ₈ DD	120000	15000	7600	4000	1900
Σ Cl ₄ DF	1200	150	300	170	200
Σ Cl ₅ DF	1100	300	290	200	160
Σ Cl ₆ DF	1600	880	400	150	110
Σ Cl ₇ DF	2100	1400	440	120	72
Cl ₈ DF	2500	1300	230	90	60
2,3,7,8-Cl ₄ DD	4.7	< 10	< 5	< 2	< 2
1,2,3,7,8-Cl ₅ DD	< 2	< 10	< 5	< 2	< 2
1,2,3,4,7,8-Cl ₆ DD	< 2	< 10	< 5	< 2	< 2
1,2,3,6,7,8-Cl ₆ DD	68	130	40	19	< 2
1,2,3,7,8,9-Cl ₆ DD	26	58	< 5	< 2	< 2
1,2,3,4,6,7,8-Cl ₇ DD	4800	2400	670	440	240
2,3,7,8-Cl ₄ DF	9.6	< 10	13	11	9.0
1,2,3,4/7,8-Cl ₅ DF	< 2	< 10	< 5	< 2	< 2
2,3,4,7,8-Cl ₅ DF	21	28	19	14	10
1,2,3,4,7,8/9-Cl ₆ DF	< 2	50	< 5	11	13
1,2,3,6,7,8-Cl ₆ DF	< 2	65	< 5	5.9	< 2
1,2,3,7,8,9-Cl ₆ DF	< 2	< 10	< 5	< 2	< 2
2,3,4,6,7,8-Cl ₆ DF	<35	53	<20	9.0	< 2
1,2,3,4,6,7,8-Cl ₇ DF	1400	730	230	74	48
1,2,3,4,7,8,9-Cl ₇ DF	21	25	< 5	< 2	< 2
TEQ	210	97	- 34	21	12

Acknowledgements

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References

- [1] Lamparski, L.L.; Nestrick, T.J.; Stenger, V.A. (1984): Presence of Chlorodibenzo Dioxins in a Sealed 1933 Sample of Dried Municipal Sewage Sludge; *Chemosphere* 13, 361-365
- [2] Hagenmaier, H.; Brunner, H.; Haag, R.; Bechthold, A. (1986): PCDDs and PCDFs in Sewage Sludge, River and Lake Sediments from South West Germany; *Chemosphere* 15, 1421-1428
- [3] Sewart, A.; Harrad, S.J.; McLachlan, M.S.; McGrath, S.P.; Jones, K.C. (1995): PCDD/Fs and Non-o-PCBs in Digested U.K. Sewage Sludge; Chemosphere **30**, 51-67
- [4] McLachlan, M.S.; Hinkel, M.; Reissinger, M.; Hippelein, M.; Kaupp, H. (1994): A Study of the Influence of Sewage Sludge Fertilization on the Concentrations of PCDD/F and PCB in Soil and Milk; *Environ. Pollution* 85/3, 337-343
- [5] Hagenmaier, H. (1990): Bundesweite Analyse von Klärschlämmen auf polchlorierte Dibenzo-p-dioxine und polychlorierte Dibenzofurane mit statistischer Stichprobenplanung und Quellenanalyse; Abschlußbericht im Auftrag des Umweltbundesamtes (UBA) Berlin, Nr.: 103 03 532
- [6] Hagenmaier, H.; Brunner, H. (1987): Isomerspecific Analysis of Pentachlorophenol and Sodium Pentachlorophenate for 2,3,7,8-Substituted PCDD and PCDF at Sub-ppb Levels; *Chemosphere* 16, 1759-1764
- [7] Rappe, C.; Kjeller, L.-O.; Andersson, R. (1989): Analyses of PCDDs and PCDFs in Sludge and Water Samples; Chemosphere 19, 13-20
- [8] Klöpffer, W.; Rippen, G.; Gihr, R.; Partscht, H.; Stoll, U.; Müller, J. (1990): Untersuchung über mögliche Quellen der polychlorierten Dibenzodioxine und Dibenzofurane in Klärschlämmen; Abschlußbericht im Auftrag des Umweltbundesamtes (UBA) Berlin, Nr.: 10303531
- [9] Näf, C.; Broman, D.; Ishaq, R.; Zebühr, Y. (1990): PCDDs and PCDFs in Water, Sludge and Air Samples from Various Levels in a Waste Water Treatment Plant with Respect to Composition Changes and Total Flux; *Chemosphere* 20, 1503-1510
- [10] Horstmann, M.; McLachlan, M.S.; Reissinger, M. (1993): Investigations of the Origin of PCDD/F in Municipal Sewage Sludge; *Chemosphere* 27, 113-120
- [11] Horstmann, M.; McLachlan, M.S. (1994): Textiles as a Source of Polychlorinated Dibenzo-p-Dioxins and Dibenzofurans (PCDD/F) in Human Skin and Sewage Sludge; *ESPR-Environ. Sci.&Pollut. Res.* 1, 15-20
- [12] Horstmann, M.; McLachlan, M.: Polychlorinated Dibenzo-p-Dioxins (PCDD) and Dibenzofurans (PCDF) in Different Textile Samples; submitted to *Chemosphere*
- [13] Kutz, F.W.; Barnes, D.G.; Bottimore, D.P.; Greim, H.; Bretthauer, E.W. (1990): The International Toxicity Equivalency (I-TEF) Method of Risk Assessment for Complex Mixtures of Dioxines and Related Compounds; *Chemosphere* 20, 751-757
- [14] Herrmann, R.; Daub, J.; Striebel, T. (1992): Charakterisierung und Analyse der Verschmutzung des Niederschlages und des Niederschlagabflußes; Abschlußbericht im Auftrag des Bundesministers für Forschung und Technologie (BMFT), Universität Karlsruhe
- [15] Kari, F.-G., Herrmann, R. (1989): Abspülung von organischen Spurenschadstoffen und Schwermetallen aus einem städtischen Einzugsgebiet: Ganglinienanalyse, Korngrößenzuordnung und Metallspeziesauftrennung; Deutsche Gewässerkundliche Mitteilungen 33, 172-183

- [16] Kern, U.; Wüst, W.; Daub, J.; Striebel, T.; Herrmann, R. (1992): Abspülverhalten von Schwermetallen und organischen Mikroschadstoffen im Straßenabfluß; gwf-Wasser-Abwasser 11, 567-574
- [17] Horstmann, M.; Kaune, A.; McLachlan, M.S.; Reissinger, M.; Hutzinger, O. (1992): Temporal Variability of PCDD/F Concentrations in Sewage Sludge; *Chemosphere* 25, 1463-1468
- [18] Schwind, K.-H. (1991): Halogenierte Dibenzo-p-dioxine und Dibenzofurane aus Verbrennungsprozessen; Dissertation, University of Bayreuth
- [19] Horstmann, M.; McLachlan, M. (1995): Polychlorierte Dibenzo-p-Dioxine und Dibenzofurane in Textilien; Melliand 1-2, 66-73
- [20] Schneider, L.(1993/4): Der umweltfreundliche Wasser-Haushalt; Hrsg.: Vereinigung Deutscher Gewässerschutz, 3. Aufl., Köln

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Date: 04-Oct-95 15:59:00 -0300

To: BSMITH@TX.NCSU.EDU (Brent Smith2)
Subject: Coming soon to your mailbox...

The document is completed and was sent to Doug Williams at EPA this week. sent our original out to have a couple of copies made and will send one to you as soon as it returns. It wouldn't hurt to go over it and take note anything that looks funny or merits change after a fresh look at it. Bet to make any changes now before it goes to printing. I suspect Doug will show it to a short list of people at EPA and then submit it for printing shortly.

I am working with Bob Pojasek on another project and he told me he was down at ATMI last week giving a presentation on P2. Something like 85 people showed up which he thought was encouraging. He said Jane Henricks told he they were very happy that EPA had taken it upon themselves to do a document for the industry and that they thought it would serve a real need.

Bob also mentioned there was some talk about a recent study where they sampled household laundry wastewater and found PCDDs and PCDFs. It seemed implicate textile finishing chemicals (bleaches? dyes?) as dioxin precursors, adn they specifically mentioned denim. Even wierder was that they found trace quantities in greywater from showers, as if the chemical had been transferred to the body and then rinsed off during showering. On course we are talking ppq here. The reference is Chemosphere vol 31, No. p. 2887-2896 and the author is Horstmann (cited in our document also?).

Heard anything about it?

At)ideas on the PCDD/F formation mechanism?

Any P2 options for reducing potential for formation?

Any chance this will become a big issue for the industry, in the same way dioxin in pulp effluent did for pulp and paper?

